

Integrating laboratory experiments and biogeographic modelling approaches to understand sensitivity to ocean warming in rare and common marine annelids

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1 **Abstract**

2 Among ectotherms, rare species are expected to have a narrower thermal niche breadth and
3 reduced acclimation capacity and thus be more vulnerable to global warming than their common
4 relatives. To assess these hypotheses, we experimentally quantified the thermal sensitivity of seven
5 common, uncommon, and rare species of temperate marine annelids of the genus *Ophryotrocha* to
6 assess those species' vulnerability to ocean warming. We measured the upper and lower limits of
7 physiological thermal tolerance, survival, and reproductive performance of each species along a
8 temperature gradient (18, 24, and 30 °C). We then combined this information to produce curves
9 of each species' fundamental thermal niche by including trait plasticity. Each thermal curve was
10 then expressed as a habitat suitability index (HSI) and projected for the Mediterranean Sea and
11 temperate Atlantic Ocean under a present day (1970-2000), mid- (2050-2059) and late- (2090-
12 2099) 21st Century scenario for two climate change scenarios (RCP2.6 and RCP8.5). Rare and
13 uncommon species showed a reduced upper thermal tolerance compared to common species, and
14 the niche breadth and acclimation capacity were comparable among groups. The simulations
15 predicted an overall increase in the HSI for all species and identified potential hotspots of HSI
16 decline for uncommon and rare species along the warm boundaries of their potential distribution,
17 though they failed to project the higher sensitivity of these species into a greater vulnerability to
18 ocean warming. In the discussion, we provide elements and caveats on the implications of our
19 results for conservation efforts.

20

21 **Keywords:** acclimation, commonness, macroecology, macrophysiology, rarity, thermal niche

22 **Introduction**

23 The recognition that organisms are unevenly distributed on Earth and that rarity is more
24 frequent than commonness has laid the foundation for the empirical investigation of the causes of
25 rarity and commonness and their consequences on large-scale ecological processes, species
26 biogeography, and biodiversity conservation (Kunin and Gaston 1997). Rarity has been an
27 important driver of species disappearance in past extinctions (Harnik et al. 2012; Calosi et al.
28 2019), and it is a widely used criterion in conservation biology and ecosystem management to
29 assess the current levels of vulnerability and extinction risk of species (IUCN 2001). Narrow
30 geographic ranges, restricted habitat specificity, small population size, and a combination of these
31 factors make rare species more at risk of decline or extinction compared with their common
32 relatives (Rabinowitz 1981; Calosi et al. 2008a). The common-rare species paradigm not only
33 defines the structure, dynamics, and function of ecosystems but is also useful for projecting the
34 current and future sensitivity of taxa and the impending reorganisation of biodiversity patterns in
35 response to ongoing climate changes (Violle et al. 2017).

36 Among the attributes of rarity, geographic range size is considered a good predictor of species
37 susceptibility to environmental perturbations (Harnik et al. 2012; Staude et al. 2020). The strict
38 correspondence between the extent of the spatial occurrence of a species and the ecological
39 requirements under which it can thrive gave origin to the *niche breadth–range size hypothesis*
40 (Brown 1984), which postulates that species with broader ecological niches can persist across a
41 wider range of environmental conditions compared to species with narrower ecological niches.
42 This hypothesis bears important implications for our understanding of the mechanisms
43 underpinning commonness and rarity (Calosi et al. 2010; Slatyer et al. 2013) and provides a
44 framework for the prediction of the redistribution of biodiversity under climate change scenarios

45 (Sunday et al. 2011, 2012). Such implications are particularly relevant for ectotherms, which
46 constitute the vast majority of species on Earth (Longhurst 2010), and for which temperature is a
47 major abiotic driver of their past, present, and future distribution *via* its pervasive effect on basic
48 physiological mechanisms, such as thermal tolerance and energy metabolism (Gagné et al. 2020;
49 Bennet et al. 2021).

50 In recent decades, considerable effort has been devoted to finding a mechanistic link
51 between the thermal niche breadth and range size of marine species to predict their distribution
52 and abundance changes under ocean warming. The inclusion of species-specific physiological
53 constraints in niche-based models has helped improve our ability to generate more accurate
54 predictions of species range shifts and species losses in the face of ocean warming (e.g., Kearney
55 and Porter 2009; Cheung et al. 2015; Reygondeau 2019). However, quantitative information on
56 the geographic occurrence and distribution of species is often insufficient for most taxa and is
57 disproportionately deficient for marine species. For many marine invertebrates, for example,
58 assessments of rarity based solely on the geographic range size is challenged by patchy
59 distributions, overdispersion, and the colonisation of microrefugia that may be difficult to detect
60 (Chapman 1999; Benkendorff and Przeslawski 2008). Accordingly, whether the applicability and
61 predictive power of the concepts of commonness and rarity also hold true in these cases has yet to
62 be tested on a wider array of taxa whose geographic distribution is not comprehensively
63 characterised. Furthermore, most niche-model studies estimate the fundamental thermal niche of
64 a species based on its mean basal thermotolerance alone, and thus these models do not account for
65 acclimation through phenotypic plasticity that can broaden the thermal requirements necessary for
66 a species to thrive (Sunday et al. 2011). Further improvements to the predictive power of niche-
67 based models can thus be obtained by integrating, within the common-rare paradigm, the

68 assumptions of those climate-based hypotheses that link a species' thermal physiology and
69 geographic range size to its acclimation ability in order to predict ectotherms' sensitivity to ocean
70 warming. Among these hypotheses, the *climatic variability hypothesis* (CVH) posits that the
71 breadth of the thermal tolerance window of a species correlates positively with its latitudinal extent
72 (Stevens 1989). An extension of the CVH asserts that widespread species or species experiencing
73 higher levels of climatic variability (e.g., common species) should have a broader physiological
74 thermal tolerance and greater acclimation capacities (i.e., *Brattstrom rule*) when compared to less
75 extensively distributed species or species colonising more stable environments (e.g., rare species)
76 (Calosi et al. 2008a,b; Bozinovic et al. 2011; Magozzi et al. 2015).

77 In this study, we tested the possibility of using the concept of commonness and rarity to
78 assess the relative levels of thermal sensitivity of marine ectotherms with differing biogeographies.
79 To do so, we used seven species of temperate, interstitial annelids of the genus *Ophryotrocha*
80 Claparède and Meczniow, 1869 (Dorvilleidae) characterised by different geographic range sizes
81 and abundances as case studies (Prevedelli et al. 2005; Simonini et al. 2009): *O. labronica* La
82 Greca and Bacci, 1962, *O. japonica* Paxton and Åkesson, 2010, *O. adherens* Paavo et al., 2000,
83 *O. puerilis* Claparède and Meczniow, 1869, *O. diadema* Åkesson, 1976, *O. hartmanni* Huth, 1933,
84 and *O. robusta* Paxton and Åkesson, 2010. We specifically assessed whether the common or rare
85 status of these species was a good predictor of their vulnerability to ocean warming. To achieve
86 our goal, we first experimentally measured the physiological thermal tolerance limits, survival,
87 and reproductive performance of the studied species along a temperature gradient, i.e., 18, 24, and
88 30 °C. We then used the generated experimental data to build the fundamental thermal niche of
89 each species following the thermal habitat suitability index (HSI) concept developed by Helaouët
90 and Beaugrand (2009), taking into account the thermal plasticity of the measured traits. Finally,

91 we spatially projected the HSI values onto a present-day (1970-2000) scenario for ocean
92 temperatures and predicted their changes under mid- (2050-2059) and late- (2090-2099) 21st
93 Century scenarios of ocean warming. Two representative concentration pathways were chosen,
94 RCP2.6 and RCP8.5, representing a ‘strong mitigation’ and ‘business-as-usual’ climate change
95 scenario, respectively (IPCC 2013). Furthermore, to identify the putative ecological mechanisms
96 underpinning the differences between common and rare species, we explored the validity of the
97 *niche breadth–range size hypothesis* in the studied species and assessed whether their
98 physiological acclimation ability was related to their level of commonness or rarity as measured
99 by the extent of their geographic range size (*Brattstrom rule*). Tests were performed on laboratory
100 strains of each species originating from subtidal populations collected in the Mediterranean region,
101 a climate change hotspot where the average sea surface temperature is predicted to increase by
102 0.93 °C to 1.24 °C over the 2010-2039 period and by 3.45 to 4.37 °C over the 2010-2099 period
103 under the RCP8.5 scenario (IPCC 2013). Ocean warming is causing the ‘tropicalization’ of the
104 southeast Mediterranean basin due to the migration of alien tropical species from the Red Sea and
105 the ‘meridionalization’ of the northern parts of the basin due to the increased occurrence of
106 indigenous thermophilic species, jeopardizing the persistence of rare, endemic, and colder-tolerant
107 species (Coll et al. 2010).

108

109 **Material and Methods**

110 *Biology and thermal habitats of the studied species*

111 The seven studied species have a similar ecology, morphology, reproductive biology, and
112 dispersal ability. They are all interstitial generalist grazers found in organic-rich benthic habitats
113 and have an average adult body length that ranges between 3 and 12 mm (*O. adherens* and *O.*

114 *puerilis*, respectively). They reproduce semicontinuously, laying clutches of eggs inside a
115 protective matrix or tube that are cared for by the parents until the eggs hatch, and the embryos
116 undergo direct development (Simonini et al. 2009, 2010). Occurrence data for these species have
117 been obtained mainly from the Northern Hemisphere. The coasts of the Shetland Islands (Scotland,
118 UK, 60° 20' 49.05'' N, 1° 14' 8.38'' W; *O. hartmanni*) and Oahu (Hawaii, USA, 21° 19' 0.23''
119 N, 157° 57' 13.04'' W; *O. labronica*) are the most northerly and southerly latitude locations,
120 respectively, where at least one of these species was found (Simonini et al. 2009). All species have
121 been recorded in the subtidal zone (first 60-90 cm below the lower tide level) of shallow marine
122 and brackish coastal waters (Simonini et al. 2010), where they showed different patterns of
123 distribution, prevalence, and abundance across thermal regions and habitats (Prevedelli et al. 2005;
124 Simonini et al. 2009). For example, the distribution range of *O. labronica*, the most ubiquitous and
125 abundant of the seven species, extends from the temperate coasts of the North Atlantic Sea to the
126 subtropical coasts of the Red Sea (Simonini et al. 2009) and comprises locations where minimum
127 and maximum annual water temperatures can range between 5 and 30 °C, respectively (Massamba-
128 N'Siala et al. 2011). In contrast, the known occurrence of the rare species *O. robusta* and *O.*
129 *diadema* is limited to Mediterranean-like climates, specifically to locations where water
130 temperatures do not exceed 27.5 °C.

131

132 *Assessment of species commonness-rarity status*

133 We identified the level of commonness or rarity of the studied species using two of the
134 three criteria suggested by Rabinowitz (1981): geographic range size and local population
135 abundance. We measured geographic range size as the extent of occurrence (EOO) (Gaston and
136 Fuller 2009), which was calculated with the tool GeoCAT (Bachman et al. 2011) using information

137 on the known global distribution of the studied species in the Northern Hemisphere (see Online
138 Resource 1, Fig. S1; Online Resource 2, Table S1). We calculated local population abundance as
139 the mean maximum number of individuals recorded in each sampled locality (Online Resource 2,
140 Table S1-S2). We then coded species as common (c), uncommon (u), or rare (r) following a
141 modified version of the ‘proportion of species’ method described by Gaston (1994). Species with
142 an EOO and local abundance higher than the third quartile of the EOO and abundance distribution
143 were considered common; *O. labronica* and *O. japonica*. Species with an EOO or local abundance
144 between the first and third quartiles of the EOO and abundance distribution were categorised as
145 uncommon: *O. adherens* and *O. puerilis*. Species with an EOO or local abundance below the first
146 quartile of the EOO and abundance distribution were considered rare: *O. hartmanni*, *O. diadema*,
147 and *O. robusta* (Online Resource 1, Fig. S2).

148

149 *Specimen maintenance and experimental design*

150 The specimens used in our study came from laboratory cultures maintained for
151 approximately ten generations (approx. 1.5 years) under a constant temperature (24 °C) and
152 salinity (35; 40 ‰ for *O. hartmanni*, as this salinity guaranteed better performance), and a 12:12
153 h light-dark regime. Founder individuals were collected in Porto Empedocle (Italy; 37°18’N;
154 13°32’E) for *O. labronica*, *O. puerilis*, *O. adherens*, *O. robusta*, and *O. diadema* and in La Spezia
155 (Italy; 44°06’N; 09°49’E) for *O. japonica* and *O. hartmanni*.

156 For each species, we produced sixty reproductive pairs (parental generation) by pairing
157 sexually mature individuals that were haphazardly collected from the laboratory cultures and by
158 isolating them in a 20 mL glass bowl under the same conditions as the cultures of origin. When
159 the first egg mass was laid and the eggs hatched (F1 generation), the parents were removed and

160 the development of the F1 individuals was followed for approximately 20 d in *O. labronica*, *O.*
161 *japonica*, *O. diadema*, and *O. robusta*, 10 d in *O. adherens*, 35 d in *O. puerilis*, and 25 d in *O.*
162 *hartmanni*. These times corresponded to the late juvenile stage that precedes the full development
163 of oocytes in the coelomatic cavity of the individuals and were used as a temporal reference to
164 randomly assign broods to one of the three acclimation temperatures: 18, 24, and 30 °C, for a total
165 of twenty broods *per* acclimation temperature, thus ensuring that exposure to the acclimation
166 temperatures started at the same developmental stage under all treatments in all species. Two of
167 the acclimation temperatures, 18 and 24 °C, were chosen within the thermal range normally
168 experienced by all the tested species in their natural environment, while 30 °C was the highest
169 temperature recorded in the site colonised by the common species *O. labronica* (Massamba-
170 N'Siala et al. 2011). Exposure to 18 and 30 °C was achieved through the increase or decrease in
171 temperature, respectively, from the culture temperature of 24 °C at a rate of 1 °C h⁻¹ (Massamba-
172 N'Siala et al. 2012) using a programmable incubator (Everlasting, AM SLIM 701 TNV,
173 Towcester, UK). At the end of this gradual pre-exposure, for each acclimation temperature, a
174 number of F1 individuals were randomly taken from the broods and assigned to one of two
175 experiments aimed at determining the physiological thermal tolerance limits (Experiment 1) and
176 measuring the survival and reproductive performance (Experiment 2) of the studied species. Each
177 of these experiments is described in detail in the following sections (see also Online Resource 1,
178 Fig. S3 for a schematic representation of the experimental design). During exposure to the
179 acclimation temperatures, the experimental replicates were isolated in 20 mL glass bowls placed
180 inside plastic containers and kept in programmable incubators (Everlasting, AM SLIM 701 TNV,
181 Towcester, UK) with a light-dark regime set at 12:12 h. Salinity (35 ‰) was kept constant
182 throughout the experiments. Artificial sea water was made by dissolving artificial sea salt (Reef

183 Crystals, Instant Ocean, St. Blacksburg, VA, USA) in distilled water. Sea water in the bowls was
184 changed every 3 d, and individuals were fed on the same day *ad libitum* with a solution of spinach
185 minced in sea water (Massamba-N'Siala et al. 2012).

186

187 *Experiment 1: Determination of physiological thermal tolerance limits*

188 Upper and lower physiological thermal tolerance limits were measured using as metric
189 *maximum* and *minimum* critical temperature limits (CT_{\max} and CT_{\min}), respectively (Lutterschmidt
190 and Hutchison 1997). CT_{\max} and CT_{\min} were defined as the temperature at which we detected
191 specific, sequential behavioural endpoints used as proxies to define the gradual physiological
192 impairment caused by exposure to suboptimal and lethal temperatures (Massamba-N'Siala et al.
193 2012, 2014). Endpoints were measured using a dynamic method (Lutterschmidt and Hutchison
194 1997), specifically by continuously observing the behaviour of individuals exposed to a constantly
195 increasing or decreasing temperature for CT_{\max} and CT_{\min} , respectively, *per* unit of time ($1\text{ }^{\circ}\text{C min}^{-1}$)
196 ¹), starting at the acclimation temperature to which individuals had been exposed for 7 d
197 (Massamba-N'Siala et al. 2012). Tests were performed on a subset of twenty individuals for CT_{\max}
198 and twenty individuals for CT_{\min} randomly taken from the broods after an exposure of 7 d to the
199 assigned acclimation temperature. We tested a maximum of five individuals per round inside a
200 borosilicate glass bowl (diam. = 3 cm, depth = 2 cm) immersed in a computer-controlled
201 recirculating ethylene glycol bath (VWR-I462-7017, Radnor, PA, USA). During each round, we
202 continuously measured the temperature of the seawater inside the glass bowl with a digital
203 thermometer (Testo 922, 2-channel Thermocouple Thermometer, Milano, Italy) equipped with a
204 precision fine wire thermocouple (T/C type K, Testo). The size of each individual was measured
205 before starting each round by counting the number of chaetigers, i.e., the segments bearing bristles

206 (Massamba-N'Siala et al. 2012). The temperatures at which the loss of locomotor control, onset
207 of spasms, and death (lethal temperature) occurred were measured on the same individual as
208 (sequential) endpoints to define CT_{max} . In contrast, the temperatures at which the onset of spasms
209 and chill coma occurred were the (sequential) endpoints measured to define CT_{min} (Massamba-
210 N'Siala et al. 2012, 2014). More specifically, loss of locomotor control was identified as the
211 occurrence of reversible arrhythmicity in locomotor activities. Lethal temperature was defined as
212 the temperature at which individuals no longer responded to poking and showed no recovery after
213 being cooled back to the corresponding acclimation temperature. The onset of spasms was defined
214 as when spasmodic, uncontrolled contraction of the entire body started. Finally, a chill coma was
215 defined as a reversible status where motionless individuals did not respond to prodding.

216 We quantified the physiological acclimation ability of the studied species following the
217 approach used by Calosi et al. (2008a). More specifically, we calculated the absolute difference
218 between the mean temperature at which an endpoint (a proxy for CT_{max} or CT_{min}) was measured
219 after acclimation at 24 and 18 °C ($\text{mean } CT_{max/min} [24 \text{ °C}] - \text{mean } CT_{max/min} [18 \text{ °C}]$).

220

221 *Experiment 2: Determination of survival and reproductive performance*

222 Individual survival and reproductive performance were measured for each species in
223 twenty pairs *per* acclimation temperature during 90 d of exposure to the assigned acclimation
224 temperature. Pairs were formed by haphazardly selecting and matching sexually matured
225 individuals taken from different broods within the same acclimation temperature to avoid
226 inbreeding (Massamba-N'Siala et al. 2011). We checked each pair daily to record the presence or
227 death of each individual and the number of eggs laid, which were counted and discarded. We
228 calculated survival as the percentage of days that an individual was alive over the total duration of

229 the experiment (90 d) and reproductive performance as the total fecundity, which was defined as
230 the total number of eggs *per* chaetiger laid by a pair (Massamba-N'Siala et al. 2011).

231

232 *Determination of a species' fundamental thermal niche*

233 Following the theoretical approach suggested by Helaouët and Beaugrand (2009; see also
234 Sunday et al. 2012; Pörtner et al. 2007), we used the experimental data on CT_{max} , CT_{min} , and total
235 fecundity to determine the shape of the fundamental thermal niche of the studied species. Curves
236 were built on four thermal ranges – for *critical tolerance*, *tolerance*, *growth*, and *reproduction* –
237 defined by the mean upper and lower limits for physiological thermal tolerance and reproduction
238 measured at each acclimation temperature and quantified in the temperature dimension as an HSI.
239 This index ranges from 0 (when the environment is unsuitable for the species) to 1 (when the
240 environment is optimal for the species). First, we bounded the thermal niche of each species using
241 the maximum mean value of the upper lethal temperature and the minimum mean temperature at
242 which chill comas occurred from each acclimation temperature. The HSI at these two designated
243 points was set at 0.01, while an HSI below or beyond this range was set as 0. Following the same
244 approach, we set the *critical tolerance* range (HSI = 0.1) at the mean value of upper lethal
245 temperature and chill coma averaged across acclimation temperatures. We then retrieved the upper
246 and lower limits of the *tolerance* range (HSI = 0.3) using the minimum mean temperature at the
247 onset of spasms (CT_{min}) and the maximum mean temperature at the loss of locomotor control,
248 respectively, measured across acclimation temperatures. We set the range for *growth* (HSI = 0.5)
249 between the maximum mean temperature at the onset of spasms (CT_{min}) and the minimum mean
250 temperature at the loss of locomotor control. We obtained the limits for the *reproduction* range
251 (HSI = 0.8) from the curve of reproductive performance derived from the data on total fecundity.

252 We assigned an HSI value of 1 to the temperature at which the maximum number of eggs was
253 estimated to be produced. Finally, we applied a shape-preserving, piecewise cubic interpolation
254 using all defined HSI values to quantify the full extent of the thermal niche from -4 to 42.7 °C
255 (mean minimum and maximum global annual sea surface temperatures) every 0.1 °C. The
256 interpolated value at a query point was based on a shape-preserving, piecewise cubic interpolation
257 of the values at thirty-five neighbouring grid points using the function *interp1* from MATLAB's
258 statistical toolbox. If a threshold point had not been quantified experimentally (e.g., total fecundity
259 for *O. puerilis* (u) and *O. hartmanni* (r)), we extrapolated the HSIs by interpolation using the other
260 existing points.

261 We calculated the breadth of the fundamental thermal niche of the species as the difference
262 between the maximum and minimum temperatures that defined each of the four ranges used to
263 build the curves of the fundamental thermal niche.

264

265 *Thermal habitat suitability maps*

266 We extracted the yearly average sea surface and bottom temperature from the IPSL-CM5-
267 MR (Institut Pierre Simon de Laplace), GFDL-ESM2 M (Geophysical Fluid Dynamics
268 Laboratory), and MPI-ESM (Max Plank Institute) Earth system models (ESMs) from the Coupled
269 Model Intercomparison Project Phase 5 (CMIP5) website for two emission scenarios, RCP2.6 and
270 RCP8.5, and for the reference (1970–2000), mid- (2050-2059) and late- (2090-2099) 21st Century
271 periods. We computed average values for each geographic cell on regular 0.5° x 0.5° grids between
272 the ESMs for each year and each decade. Then, using the fundamental niche previously computed,
273 we retrieved the corresponding HSI values averaged for each grid cell using sea surface and bottom
274 temperatures. In this way, we calculated HSI values for each species in the North Atlantic and

275 Mediterranean regions at the different periods for the two RCPs. Since the studied species are
276 associated with coastal environments, HSI values located in the open ocean domain were set to 0
277 following the biogeographic raster of Reygondeau et al. (2013). In addition, to better visualise the
278 changes in HSI under the two RCPs, we plotted maps to show the mathematical difference between
279 the HSI values projected in the future scenarios and those calculated for the reference period.
280 Following the approach described above, we also calculated HSI values for each species based on
281 real temperature data for the North Atlantic and Mediterranean regions using the World Ocean
282 Atlas 2018 database (Locarnini et al. 2019) and for the Mediterranean Sea using the
283 MEDAR/MEDATLAS database (Fichaut et al. 2003). These maps showed a good match with
284 those obtained for the reference period (1970–2000) used in our projections (Online Resource 3,
285 Appendix 1; Online Resource 1, Fig. S4-S5; Online Resource 2, Table S3).

286

287 *Statistical analyses*

288 We tested the effects of species status (*Status*; three levels: c, u, r), species identity
289 (*Species(Status)*; seven levels, nested in *Status*), acclimation temperature (*T*; two levels: 18, 24
290 °C), and their interactions on CT_{max} and CT_{min} with generalised linear models (GLMs; family
291 Gaussian, identity link), using body size as a covariate. This analysis was performed for only two
292 of the three acclimation temperatures because none of the rare species and only one uncommon
293 species, *O. adherens*, survived a 7 d exposure to 30 °C. We then performed a second analysis with
294 GLMs (family Gaussian, identity link), using body size as a covariate, to test for the effect of
295 species identity (*Species*; three levels), acclimation temperature (*T*; three levels: 18, 24, 30 °C),
296 and their interaction on the CT_{max} and CT_{min} of the three species that survived the exposure to 30

297 °C: *O. labronica* (c), *O. japonica* (c), and *O. adherens* (u). For this latter analysis, there was no
298 replication for the species *Status*.

299 Due to technical problems, all the specimens of *O. puerilis* (u) and *O. hartmanni* (r) died before
300 being paired for the measurement of survival and reproductive performance. Consequently, we
301 used GLMs to assess the effect of species identity (*Species*; five levels), acclimation temperature
302 (*T*; three levels: 18, 24, 30 °C), and their interactions on survival (family quasipoisson, log link)
303 and total fecundity (family Gaussian, identity link) for the remaining species, as there was no
304 replication for species *Status*.

305 Post hoc tests were performed with Tukey's test using the R package *multcomp*. Since multiple
306 endpoints, used to define physiological traits (CT_{max} or CT_{min}), or multiple life-history traits
307 (survival and total fecundity) were measured on the same individual, we applied the Benjamini–
308 Hochberg correction for false positives to all *P*-values.

309 Preliminary analyses showed that none of the measured traits exhibited a phylogenetic signal ($\lambda \neq$
310 0, $P < 0.05$), which was estimated using Pagel's λ using the 'phylosig' function in the R package
311 *phytool* (for details, see Online Resource 3, Appendix 2; Online Resource 1, Fig. S6; Online
312 Resource 2, Table S4). Accordingly, we performed final analyses without applying any
313 phylogenetic correction.

314 We investigated the relationship between the physiological acclimation ability and
315 geographic range size for each proxy of CT_{max} and CT_{min} using a phylogenetic generalised least
316 squares multiple regression with the R package *ape*, using body size as a covariate. Using the same
317 technique, we assessed the relationship between niche breadth and geographic range size by fitting
318 linear models for the breadth of the fundamental thermal niche as a function of the EOO for each
319 of the four performance ranges previously identified: *critical tolerance*, *tolerance*, *growth*, and

320 *reproduction*. Assumptions of normality and homogeneity of variance were checked on the
321 residuals with the Shapiro–Wilk test and Levene test, respectively, and were met for all the
322 regression models.

323 We performed all statistical analyses using R software, version 4.0.0 (RStudio Team 2020; see
324 Online Resource 4 for R scripts).

325

326 **Results**

327 *Physiological upper thermal tolerance limits*

328 Mean CT_{max} (\pm SE) ranged between 33.19 ± 0.03 and 41.82 ± 0.02 °C in the group of
329 common species (*O. labronica* and *O. japonica*), between 27.28 ± 0.79 and 35.70 ± 0.05 °C in the
330 group of uncommon species (*O. adherens* and *O. puerilis*), and between 29.33 ± 0.04 and $36.58 \pm$
331 0.04 °C in the group of rare species (*O. hartmanni*, *O. diadema*, and *O. robusta*) (Fig. 1a-c).
332 Common species had a significantly higher mean CT_{max} than their uncommon and rare congeners
333 both at 18 and 24 °C and for all endpoints (*Status* by $T_{(18,24\text{ °C})}$: maximum deviance = 182.48, $P <$
334 0.001 ; Table 1a-c, Fig. 1a-c. See Online Resource 2, Table S5 for pairwise comparisons).
335 Uncommon species showed a slightly but significantly lower mean CT_{max} compared to the rare
336 species when measured as the loss of locomotor control (Fig. 1a) and lethal temperature (Fig. 1c)
337 following exposure to 18 °C. In all other cases, uncommon and rare species showed comparable
338 mean CT_{max} (Fig. 1a-c).

339 We found significant interspecific differences for all the endpoints used to define CT_{max}
340 (*Species(Status)* by $T_{(18,24\text{ °C})}$: maximum deviance = 144.67, $P < 0.001$; Table 1a-c; Fig. 2a-c).
341 Mean CT_{max} generally increased with increasing acclimation temperatures for all species when
342 measured as the loss of locomotor control (Fig. 2a) and for all species except *O. japonica* (c) when

343 measured as the onset of spasms (Fig. 2b). This trend was less consistent in the measurement of
344 lethal temperature, for which three out of the five rare-uncommon species showed comparable
345 levels of mean CT_{max} at 18 and 24 °C (Fig. 2c). Mean CT_{max} was the highest in *O. labronica* (c),
346 followed by *O. japonica* (c) (Fig. 2a-c. See Online Resource 2, Table S6 for pairwise comparisons
347 and Table S7 for mean raw values). In general, *O. diadema* (r) and *O. robusta* (r) showed lower
348 mean CT_{max} than *O. japonica* (c). *Ophryotrocha hartmanni* (r) and *O. puerilis* (u) stood out for
349 having the lowest mean CT_{max} of all species (Fig. 2a-c). Finally, *O. adherens* (u) had among the
350 lowest mean CT_{max} when measured as the loss of locomotor control (Fig. 2a) and had mean CT_{max}
351 between those measured for *O. japonica* (c) and *O. robusta* (r) when measured as the onset of
352 spasms and the lethal temperature (Fig. 2a-c).

353 The lowest and highest mean (\pm SE) CT_{max} ranged between 34.34 ± 0.14 °C (*O. adherens*
354 (u)) and 42.66 ± 0.13 °C (*O. labronica* (c)), respectively, following exposure to 30 °C (Online
355 Resource 1, Fig. S7; Online Resource 2, Table S7 for mean raw values). Analyses performed on
356 the species that survived the exposure to 30 °C confirmed the previously observed patterns: *O.*
357 *labronica* (c) and *O. japonica* (c) showed higher mean CT_{max} than *O. adherens* (u) at all
358 acclimation temperatures for the loss of locomotor control, onset of spasms (*Species* by $T_{(18,24,30}$
359 °C): maximum deviance = 99.07, $P < 0.001$; Table 2a-b. See Online Resource 2, Table S8 for
360 pairwise comparisons) and lethal temperature (*Species*: deviance = 68.63, $P < 0.001$; Table 2c;
361 Online Resource 1, Fig. S8). Finally, mean CT_{max} also increased with increasing acclimation
362 temperature when 30 °C was included in the analysis. The only exception to this pattern was found
363 in the loss of locomotor control in *O. japonica* (c), for which no significant differences were
364 detected between CT_{max} measured at 24 and 30 °C (Online Resource 1, Fig. S7).

365

366 *Physiological lower thermal tolerance limits*

367 Mean (\pm SE) CT_{min} ranged between -1.69 ± 0.02 and 8.08 ± 0.02 °C in the group of common
368 species, -2.07 ± 0.05 and 8.53 ± 0.07 °C in the group of uncommon species, and -2.40 ± 0.01 and
369 8.20 ± 0.02 °C in the group of rare species (Fig. 1d, e). The common species had slightly lower
370 mean CT_{min} compared to uncommon species when measured as the onset of spasms and had
371 slightly lower mean CT_{min} compared to both uncommon and rare species when measured as chill
372 coma following exposure to 24 °C. No significant difference in mean CT_{min} was found between
373 these three groups in any other cases (*Status* by $T_{(18-24\text{ °C})}$: maximum deviance = 169.97, $P = 0.01$;
374 Table 1d,e; Fig. 1d,e. See Online Resource 2, Table S5 for pairwise comparisons).

375 CT_{min} varied differently among species depending on the acclimation temperature
376 (*Species(Status)* by $T_{(18,24\text{ °C})}$: deviance = 117.91, $P < 0.001$; Table 1d,e; Fig. 2d,e). Mean CT_{min}
377 generally increased with increasing acclimation temperatures in all species. An exception to this
378 trend was observed in *O. labronica* (c), which showed comparable mean CT_{min} at 18 and 24 °C
379 when measured as the chill coma (Fig. 2e). In general, *O. adherens* (u) and *O. hartmanni* (r)
380 showed the highest mean CT_{min}, while *O. puerilis* (u) showed the lowest mean CT_{min} (Fig. 2d,e;
381 Online Resource 2, Table S6 for pairwise comparisons and Table S7 for mean raw values). The
382 remaining species showed intermediate mean CT_{min} that ranked differently across acclimation
383 temperatures depending on the endpoint considered (Fig. 2d,e).

384 The lowest and highest mean (\pm SE) CT_{min} ranged between -0.63 ± 0.23 °C (*O. japonica*
385 (c)) and 12.45 ± 0.23 °C (*O. adherens* (u)), respectively, following exposure to 30 °C (Online
386 Resource 1, Fig. S7; Online Resource 2, Table S7 for mean raw values). *Ophryotrocha adherens*
387 (u) also had the highest mean CT_{min} following exposure to 30 °C (*Species* by $T_{(18,24,30\text{ °C})}$: maximum
388 deviance = 72.94, $P < 0.001$; Table 2d,e. See Online Resource 2, Table S8 for pairwise

389 comparisons). Finally, mean CT_{\min} generally decreased with increasing acclimation temperatures
390 in all three species and for all endpoints when exposure to 30 °C was included in the analyses
391 (Online Resource 1, Fig. S7).

392

393 *Survival and reproductive performance*

394 At 18 °C, uncommon and rare species showed a comparable mean survival to *O. japonica*
395 (c) (100% at all acclimation temperatures) and a higher mean (\pm SE) survival compared to *O.*
396 *labronica* (c) ($76.78 \pm 6.41\%$) (*Species* by $T_{(18,24,30\text{ °C})}$: deviance = 509.30, $P < 0.001$; Fig. 3a,
397 Table 3a. See Online Resource 2, Table S9 for mean raw values and Tables S10 for pairwise
398 comparisons). At 24 °C, the mean survival of *O. diadema* (r) and *O. robusta* (r) was comparable
399 to that measured in *O. japonica* (c), as well as comparable to that measured at 18 °C, while the
400 mean survival of *O. adherens* (u) was lower than that measured at 18 °C ($49.79 \pm 2.28\%$).
401 Survival of *O. labronica* (c) at 24 °C showed mean values comparable to those measured at 18
402 °C and were intermediate among those measured for *O. adherens* (u) and all the other species. At
403 30 °C, *O. japonica* (c) and *O. labronica* (c) ($97.89 \pm 1.77\%$) had comparable mean survival, and
404 both common species had a significantly higher mean survival than uncommon and rare species;
405 mean values for survival in the latter two groups ranged between 14.67 ± 2.57 and $11.32 \pm$
406 1.53% in *O. robusta* (r) and *O. adherens* (u), respectively (Fig. 3a).

407 *Ophryotrocha robusta* (r) was the most fecund species at 18 and 24 °C (60.39 ± 3.93 and
408 67.57 ± 10.16 eggs chaetigers⁻¹, respectively), followed by *O. adherens* (u) and *O. labronica* (c)
409 at 18 °C and by *O. adherens* (u), *O. labronica* (c), and *O. japonica* (c) at 24 °C (*Species* by $T_{(18,24,30$
410 $\text{ °C})}$: deviance = 26027, $P < 0.001$; Fig. 3b, Table 3b. See also Online Resource 2, Table S9 for mean
411 raw values and Tables S10 for pairwise comparisons). The least fecund species were *O. japonica*

412 (c) and *O. diadema* (r) at 18 °C and *O. diadema* (r) at 24 °C. At 30 °C, *O. labronica* (c) showed a
413 mean total fecundity that was higher than *O. japonica* (c) (40.78 ± 6.75 and 27.79 ± 2.96 eggs
414 chaetigers⁻¹, respectively), and both common species were on average more fecund than the
415 uncommon and rare species. More specifically, *O. diadema* (r) and *O. robusta* (r) did not
416 reproduce, while *O. adherens* (u) had an extremely low mean total fecundity (0.56 ± 0.20 eggs
417 chaetigers⁻¹) (Fig. 3b).

418

419 *Acclimation ability, thermal niche breadth, and range size*

420 Physiological acclimation ability was not related to the geographic range size of the studied
421 species for all the analysed endpoints (Online Resource2, Table S11). Similarly, we found no
422 significant relationship between the breadth of the fundamental thermal niche and range size in all
423 the performance ranges considered (Online Resource 2, Table S12).

424

425 *Fundamental thermal niches and thermal habitat suitability*

426 The fundamental thermal niche of the studied species was represented by a unimodal, left-
427 skewed curve, where the thermal habitat suitability (HSI) increases with increasing temperatures
428 until it reaches a peak at the optimal temperature for reproduction and steeply descends thereafter
429 (Fig. 4). Common species extended their HSI towards higher temperatures for all functions and
430 maximised their reproductive performance under warmer conditions (25-27 °C) compared with
431 uncommon and rare species (21-23 °C), thereby intensifying the asymmetry of their curves (Fig.
432 4). In contrast, the left part of the curves unfolded within a narrower range of temperatures and
433 had no recognisable patterns in variation between species.

434 Projections based on the fundamental thermal niche of the studied species showed that HSI
435 values higher than 0.8 ranged between the 30th and 45th parallel latitudes when mapped in the
436 reference period (Online Resource 1, Fig. S9-S10). An overall increase in the HSI was observed
437 for all species under both RCP scenarios and was much more marked poleward, especially for *O.*
438 *labronica* (c) and *O. robusta* (r) (Fig. 5 and Online Resource 1, Fig. S11). In both the reference
439 and future scenarios, uncommon and rare species tended to possess higher HSI values (0.9-1) in
440 many more cells than their common relatives, particularly at the lowest latitude of their potential
441 distribution (Fig. 5 and Online Resource 1, Figs. S11). The HSI values decreased locally under
442 RCP 8.5 in some of the uncommon and rare species, specifically along the coasts of the Levantine
443 Sea (*O. hartmanni* (r)), the Sirte Gulf (Libya) (*O. puerilis* (u), *O. adherens* (u), and *O. hartmanni*
444 (r)), the Atlantic coasts of North Africa (*O. hartmanni* (r) and *O. adherens* (u)), and the southern
445 coasts of the USA (*O. puerilis* (u) and *O. hartmanni* (r)) (Fig. 5 and Online Resource 1, Fig. S11).

446

447 **Discussion**

448 Our study sheds light on a number of fundamental physiological and ecological
449 mechanisms explaining the commonness and rarity of temperate marine ectotherms, showing that
450 uncommon and rare species are more physiologically sensitive to elevated temperatures than
451 common species due to their lower physiological heat tolerance. Our simulations also suggest that
452 uncommon and rare species may be more prone to declines along the warm boundaries of their
453 potential distribution. However, the simulations fail to project these species' higher thermal
454 sensitivity into their overall greater vulnerability to ocean warming. Our results are consistent with
455 the known biogeography of the studied species and support the use of the dichotomous
456 commonness and rarity concept to assess macrophysiological patterns of thermal sensitivity in

457 ectotherm species. In addition, our results indicate that the predictive power of this conceptual
458 framework is less accurate when dealing with intermediate levels of rarity, for example, the
459 uncommon *versus* rare species in this case.

460

461 *Upper thermal tolerance: a driving factor in defining biogeographic differences*

462 We find strong support for the idea that common species possess a greater physiological
463 heat tolerance than their uncommon and rare congeners at all acclimation temperatures and for all
464 lethal and sublethal endpoints measured. In addition, common species outperformed uncommon
465 and rare species at the highest acclimation temperature, showing higher survival and reproductive
466 capacities. Similar results were obtained by Thibault et al. (2020), who found that *O. japonica* (c)
467 was able to persist for two generations at an elevated temperature (28 °C) thanks to the beneficial
468 adjustment of its energy metabolism, as characterised using a targeted metabolomics approach,
469 while *O. robusta* (r) faced a lethal increase in energy requirements before the first generation of
470 viable offspring could be produced. In contrast, no clear physiological or ecological patterns linked
471 to the biogeographical status of the studied species were found when cold tolerance was considered
472 or when survival and reproductive performance were measured and compared within the 18 to 24
473 °C range.

474 Our observations provide a mechanistic explanation of the known global, regional, and
475 local geographic distribution of the studied species. For example, the greatest heat tolerance of *O.*
476 *labronica* (c) accounts for its broadest geographic range size and its ability to persist in tidal and
477 subtidal waters where temperatures as high as 30 °C have been recorded (Massamba-N'Siala et al.
478 2011). Not surprisingly, *O. labronica* (c) has been the only species to date found in subtropical
479 waters, specifically along the coasts of the Northern Red Sea (Simonini et al. 2009). *Ophryotrocha*

480 *japonica* (c), a species originally from the North Pacific coasts, is the other most heat tolerant
481 species. Since it was first recorded in the Mediterranean Sea, this species was able to colonise the
482 same thermal regions as *O. labronica* (c) and now has the second largest range size extent among
483 the studied species (Simonini et al. 2009). In contrast, *O. hartmanni* (r), the least heat tolerant
484 species, is characterised by very low densities in subtidal waters (1-7 indiv.; Online Resources 2,
485 Table S1) but is among the most abundant macrobenthic species in the sediments of fish farms on
486 the North Atlantic coasts (Pereira et al. 2004), suggesting its preference for colder waters.
487 Similarly, *O. puerilis* (u) is among the most physiologically sensitive species to elevated
488 temperatures and it capitalizes on thermal habitats that better match its preference for colder
489 temperatures. For example, by colonising organic-rich sediments at higher depths (Taboada et al.
490 2017) or by reaching higher local densities during the coldest months of the year (Prevedelli et al.
491 2005).

492 Of course, several factors other than temperature can drive the spatial and temporal
493 distribution of marine species (Gagné et al. 2020). Different sensitivities to other abiotic factors
494 (such as salinity, oxygen, pH), biotic interactions, capacity and limitation to dispersal, habitat
495 preference, and colonisation history can prevent a species from fully realising its fundamental
496 thermal niche, contributing to a mismatch between the physiological thermal tolerance and
497 distribution range of a species (i.e., realised niche, Pulliam 2000; Arribas et al. 2012; Sánchez-
498 Fernández et al. 2012). For example, *O. diadema* (r), a species initially reported in the Pacific
499 coasts, and *O. robusta* (r), a species endemic to the Mediterranean Sea (Simonini et al. 2009),
500 display a greater physiological heat tolerance among rare species, but their known distribution is
501 limited to only two and eight known localities, respectively (Online Resource 1, Fig. S1). For these
502 species, upper thermal tolerance limits are expected to be major determinants of their colonisation

503 success in a new place. However, their recent colonisation history (for *O. diadema*) or the limited
504 chance of dispersion of the species of the genus *Ophryotrocha*, which usually passively occurs
505 through the ballast water of ships (Simonini et al. 2009), coupled with the absence of thermal
506 microrefugia along the dispersal trajectories due to the patchy occurrence of suitable habitats, may
507 play an important role in determining the rarity level of these annelid species.

508

509 *Macrophysiological and macroecological patterns for common and rare species*

510 Over the past decades, an increasing number of studies have attempted to use thermal
511 tolerance limits to search for macrophysiological and macroecological patterns that could inform
512 species vulnerability to climate change (e.g., Calosi et al. 2008a; Calosi et al. 2010; Bozinovic et
513 al. 2011; Sunday et al. 2011, 2012). Most documented patterns have produced hypotheses linking
514 the organismal thermal physiology and geographic range size. Among these hypotheses, the
515 *climatic variability hypothesis* (CVH) and its corollaries (e.g., *Brattstrom rule*) predict a positive
516 association between a species' geographic range size, physiological thermal tolerance, and
517 acclimation capacities (Bozinovic et al. 2011; Magozzi and Calosi 2015). In diving beetles of the
518 genus *Deronectes*, for example, heat and cold thermal tolerance were positively correlated with
519 range size, and restricted species were foreseen to be most at risk from global warming due to their
520 reduced acclimation capacities (Calosi et al. 2008a; Calosi et al. 2010). Our results provide partial
521 support for the CVH: a broader range size was associated with a greater physiological heat
522 tolerance in the two common species, but we report no quantitative link between the species range
523 extent and heat tolerance within the group of uncommon and rare species. In addition, no
524 relationship appears to exist between a species' range size and its physiological acclimation ability,
525 and upper and lower thermal tolerance limits increase and decrease with increasing acclimation

526 temperatures, respectively, in all species. The fact that we used strains originating from subtidal
527 temperate populations that are adapted to seasonal thermal changes may explain this result.

528 The existence of a relationship between a species' thermal tolerance breadth and its range
529 size is another mechanism suggested to explain differences in the distribution of common and rare
530 species (Brown 1984). The *niche breadth–range size hypothesis* has recently received empirical
531 validation (Slayter et al. 2013; Stuart-Smith et al. 2017), and numerous large-scale studies have
532 demonstrated the tight relationship existing between the width of the physiological thermal
533 tolerance window and the extent of geographic range boundaries across latitudinal gradients (e.g.,
534 Calosi et al. 2010; Compton et al. 2007; Sunday et al. 2011). In our study, the breadth of the
535 fundamental thermal niche failed to predict the extent of the range size. Our result contributes to
536 the ongoing debate around whether the *niche breadth–range size hypothesis* is universally valid
537 (e.g., Hirst et al. 2017). A mismatch between the scale at which niche breadth and range size are
538 measured could explain why our data provided no support for this hypothesis (Kambach et al.
539 2019).

540 The comparable or higher thermal sensitivity that uncommon species show compared to
541 rare species may (at least in part) help explain why our results provide weak or no support for the
542 tested hypotheses (*CVH, niche breadth–range size hypothesis*). On one hand, this highlights the
543 limitations of the approach used to categorise species into discrete biogeographical groups within
544 a *continuum* of forms, commonness and rarity, which are relative concepts. Rare-common species
545 cut-off points can, in fact, change depending on the level of rarity of other congeneric species, the
546 spatial and temporal scale of investigation, or following the inclusion of other criteria (Gaston
547 1994; Flather and Sieg 2007), potentially leading to different conclusions on the relative
548 differences in thermal sensitivity between uncommon and rare species. Currently, there are only a

549 few continuous indices of rarity, and the most used indices (in the terrestrial realm) are based on
550 exhaustive distributions (mostly from expert range maps) that are not available for marine
551 invertebrates. In fact, for this group of metazoans, gathering information on the criteria used for
552 categorising common and rare species (e.g., range size) is particularly challenging. The use of
553 more flexible categorisation approaches could help increase the accuracy with which closely
554 related species are arranged into intermediate levels of species rarity. This could be achieved by
555 including metrics that are generally considered less rigorous in predicting species distribution, such
556 species prevalence, i.e., the presence/absence ratio to the total number of sites collected (Jiménez-
557 Valverde et al. 2009). For example, the inclusion of information on prevalence among the criteria
558 used to define the common-rare status of our studied species would have made the distinction less
559 marked between *O. adherens* (u), *O. diadema* (r), and *O. robusta* (r), all of which had a prevalence
560 lower than or close to the first quartile (0.08) that defined the uncommon-rare species cut-off point
561 for this trait (Online Resource 2, Table S2). These observations were confirmed by the occurrence
562 data obtained for the studied species in the Mediterranean Sea, a region subjected to a more
563 extensive and systematic decadal collection effort and thus less biased by unbalanced sampling
564 compared to global distribution records (Simonini et al. 2009). On the other hand, we cannot
565 discard the possibility that species thermal physiology alone cannot provide a comprehensive
566 explanation of intermediate levels of rarity. As discussed above, temperature is not the only factor
567 influencing the geographic distribution of uncommon and rare species. The role of other abiotic
568 factors (e.g., Sunday et al. 2014), interspecific interactions (e.g., Godsoe and Harmon 2012), and
569 the number of dispersal opportunities (e.g., Grantham et al. 2003; Arribas et al. 2012) is worth
570 considering in future investigations of the mechanisms explaining rarity.
571

572 *Habitat suitability shift in common and rare species under ocean warming: conclusions and*
573 *caveats*

574 The unimodal, left-skewed curves of the fundamental thermal niche of the studied species
575 resembled the thermal performance curves used to describe the thermal sensitivity of biological
576 performances in ectotherms (Schulte et al. 2011). The level of asymmetry of these curves, as well
577 as the position of the thermal *optimum* relative to the mean of the environmental temperature, have
578 important implications for predicting species vulnerability to ocean warming. In fact, species
579 exposed to thermal conditions that are close to or above their thermal *optimum* for performance
580 are expected to have narrower or no safety margins to cope with increasing temperatures and,
581 consequently, be more at risk of population declines under ocean warming (Deutsch et al. 2008).
582 In our study, this prediction should have been relevant for *O. puerilis* (u), *O. adherens* (u), and *O.*
583 *hartmanni* (r) due to the closer proximity of their thermal *optimum* to their upper thermal limits
584 for reproduction, as confirmed (see below) by our simulations.

585 The spatial projections of the HSI captured the overall latitudinal range where this group
586 of annelids is found. This correspondence is better represented when the fundamental niche is
587 defined by the thermal range for reproduction ($0.8 < \text{HSI} < 1$) rather than the range for
588 physiological thermal tolerance ($0.1 < \text{HSI} < 0.5$). The former is in fact expected to provide a more
589 accurate approximation of the conditions that shape the ecological success of a species (Helaouët
590 and Beaugrand 2009). In both ways, *O. labronica* (c) is the only species shown to be a thermal-
591 range conformer: it exhibits a close match between its latitudinal range and its thermal tolerance
592 (Sunday et al. 2012). The idea of the presence of a ‘hyper’ performing species within the annelid
593 genus *Ophryotrocha* echoed what was observed in the widespread diving beetle *Deronectes latus*
594 Stephens, 1829 (Calosi et al. 2008a; Calosi et al. 2010) and the gammarid *Gammarus duebeni*

595 Lilljeborg, 1852 (Gaston and Spicer 2001). Both *D. latus* and *G. duebeni* are the most widespread
596 and physiologically tolerant species of their genus, supporting Gaston and Spicer's (2001) idea of
597 the existence of species that are jacks-of-all-trades and masters-of-all, further confirmed by
598 observations on the thermal metabolic performance and tolerance plasticity of the invasive
599 decapod Asian shrimp *Palaemon macrodactylus* Rathbun, 1902 (Magozzi and Calosi 2015).
600 Conversely, all of the other species investigated in our study did not meet their potential
601 distribution range. The mismatch between the geographic range of a species and its potentially
602 occupied range based on its thermal physiology can reduce the accuracy with which projections
603 that are based only on temperature can generate predictions on range shifts (Aubry et al. 2017).

604 Our simulations foresee two trends of HSI shifts extensively documented in temperate
605 marine ectotherms in response to ocean warming: (i) range expansion towards higher latitudes and
606 (ii) range contraction at lower latitudes (Cheung et al. 2009; Morley et al. 2018). These patterns
607 varied depending on the common-rare status of the studied species. In the reference decades used
608 in this study, uncommon and rare species commonly lived closer to their thermal *optimum* for
609 reproduction and at many more sites than common species. This pattern is projected to intensify
610 at the end of the 21st Century, particularly at the lowest latitudes of the Atlantic and Mediterranean
611 regions, where uncommon and rare species are projected to live at their optimal temperatures for
612 performance in most localities, or above it in some regions at the warm edges of their potential
613 distribution under RCP8.5. Here, we identified hotspots of potential population decline for *O.*
614 *adherens* (u), *O. puerilis* (u), and *O. hartmanni* (r), the three species that showed narrower upper
615 safety margins for niche adjustment. Some of these areas, such as the Levantine Sea, are already
616 known for their high occurrence of extinction events of endemic species (Rilov 2016). It must be
617 noted, however, that such a scenario seems to be potentially relevant only for *O. adherens*, the

618 only species recorded in this area. Indeed, our projections suggest that ocean warming may have
619 an overall beneficial effect on the persistence of these annelid species, as indicated by the increased
620 HSI predicted for all the species across their potential distribution. Such predictions may be more
621 optimistic, as they were built on the thermal responses of strains subjected to laboratory
622 conditioning over multiple generations. Exposure to acute, severe conditions in subtidal habitats
623 can in fact exert a positive influence on an organism's thermal performance *via* the activation of
624 mechanisms regulating heat stress responses (Giomi et al., 2019). Consequently, we cannot rule
625 out the possibility that our studied species may be able to withstand warming waters more
626 efficiently in their natural habitat compared to an experimental setting. Nevertheless, the
627 magnitude of these considerations must be carefully examined in light of the limitations of the
628 methodological approach used and the ecology of the studied species. First, our simulations are
629 based on scenarios of ocean warming that rely on yearly averaged temperatures that do not
630 accurately predict the thermal conditions on which coastal species set their thermal *optima*
631 (Boersma et al. 2016; Stuart-Smith et al. 2017). This is particularly true in subtidal habitats, where
632 temperature can vary greatly at a short temporal scale and where the expected increase in the
633 frequency, intensity, and duration of extreme heat events may pose an additional challenge to the
634 persistence of less tolerant species. This highlights the importance of capturing the heterogeneity
635 of thermal conditions at a finer scale to identify the presence of microrefugia for thermally
636 sensitive species (Potter et al. 2013). However, temperature records that are more relevant to the
637 scale or habitat under investigation are often unavailable or not easily accessible to the scientific
638 community. Second, poorer dispersers are expected to have a lower probability of reaching new
639 suitable habitats and expanding their geographic range towards higher latitudes (Thomas et al.
640 2004). Species of the genus *Ophryotrocha* can actively disperse by drifting, though presumably

641 not for long distances given the presence of direct developing larvae. Scant dispersal abilities may
642 limit the potential poleward expansion projected for these annelid species, especially for
643 populations found in semi-enclosed basins such as the Mediterranean Sea. If tracking more
644 favourable thermal conditions is not an option, climate-driven local declines or extinction events
645 may be avoided through plastic or adaptive behavioural and/or physiological responses. Subtidal
646 organisms are known to possess physiological adaptations to cope with thermally stressful events,
647 either through acclimatisation or the evolution of higher thermal tolerance maxima and thermal
648 performance breadth (Tomanek 2010; da Silva et al. 2019). In our simulations, the inclusion of
649 species' physiological and reproductive thermal plasticity contributed to an expanded projected
650 thermal habitat suitability for all species but did not prevent the occurrence of HSI decreases in
651 some of the uncommon and rare species. This result suggests that the acclimation potential of a
652 species does not guarantee its persistence in a climate-change context, as previously reported by
653 other authors (e.g., Gunderson and Stillman 2015; Magozzi and Calosi 2015).

654 In conclusion, we provide general support for the possible use of the concepts of commonness and
655 rarity to predict the susceptibility to rapid environmental changes in temperate benthic
656 invertebrates, particularly in marine annelids, a ubiquitous and diverse group of metazoans that are
657 unrepresented in macroecological and macrophysiological studies. By combining experimental
658 observations and macroecological-physiological approaches, we found a greater thermal
659 sensitivity in uncommon-rare species compared to their common relatives, although information
660 on those species' geographic distribution was scant, patchy, or discontinuous. However, while the
661 experimental observations and simulations based on the common species' thermal performance
662 was in line with theoretical expectations, predictions on the fate of uncommon and rare temperate
663 species under ocean warming scenarios were less accurate and more susceptible to methodological

664 limitations. This highlights the need to develop an even more rigorous definition of the concept of
665 rarity and its intermediate levels for integration with scenarios of climate change that are more
666 relevant for the species under investigation if we want to more accurately predict the vulnerability
667 of 'elusive' taxa to climate change.

668 **Declarations**

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676 **Ethical approval:** All applicable institutional and/or national guidelines for the care and use of
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678 **Consent to participate:** Not applicable.

679 **Consent for publication:** Not applicable.

680 **Availability of data and material:** The datasets analysed during the current study were deposited
681 in Dryad with the doi: <https://doi.org/10.5061/dryad.80gb5mksq>.

682 **Code availability:** see Supplementary Information “MassambaNSiala_et_al_ESM4_Rscripts”.

683 **Author contributions:** The experimental design has been conceived and planned by GMN.
684 Experimental measurements were carried out by GMN with the support of RS and DP. GMN
685 conducted statistical analyses with advice from RS and PC. Data modelling was conceived by
686 GMC, GR, WWC, PC and performed by GR. GMN wrote the first draft of this manuscript
687 supported by PC. All authors contributed to the final version of the manuscript.

688

689

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Table 1. Results of the analysis of deviance for the effect of species *Status* (common, uncommon, rare), species identity (*Species(Status)*) and acclimation temperature (*T*: 18 and 24 °C), and their interactions on the upper (a-c) and lower (d-e) physiological thermal tolerance limits (CT_{max} and CT_{min}, respectively) of the species of the genus *Ophryotrocha*. Degrees of freedom (Df), deviance residuals (Dev. Res.), degrees of freedom residuals (Df Res.), deviance (Dev.), and *P*-value with Benjamini-Hochberg correction for false positives (*P*(BH)) are provided

CT _{max}						CT _{min}					
a) Loss of locomotor control						d) Onset of spasms					
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)	Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)
Null			245	2154.10		Null			245	776.64	
Body size	1	423.98	244	1730.12	8.25E-16	Body size	1	37.04	244	739.60	5.50E-16
Status	2	733.80	242	996.32	8.25E-16	Status	2	7.55	242	732.05	0.001
T	1	553.52	241	442.80	8.25E-16	T	1	362.48	241	369.57	5.50E-16
Species(Status)	4	229.63	237	213.17	8.25E-16	Species(Status)	4	194.45	237	175.12	5.50E-16
Status*T	2	30.69	235	182.48	5.70E-11	Status*T	2	5.15	235	169.97	0.01
Species(Status)*T	4	37.80	231	144.67	7.31E-12	Species(Status)*T	4	52.07	231	117.91	5.50E-16
b) Onset of spasms						e) Chill coma					
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)	Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)
Null			245	1866.57		Null			245	728.72	
Body size	1	763.69	244	1102.88	8.25E-16	Body size	1	139.63	244	589.09	3.67E-16
Status	2	730.77	242	372.11	8.25E-16	Status	2	27.03	242	562.06	3.67E-16
T	1	80.46	241	291.65	8.25E-16	T	1	87.33	241	474.73	3.67E-16
Species(Status)	4	166.39	237	125.26	8.25E-16	Species(Status)	4	301.06	237	173.68	3.67E-16
Status*T	2	4.49	235	120.76	0.01	Status*T	2	25.60	235	148.07	3.67E-16
Species(Status)*T	4	24.51	231	96.26	1.55E-11	Species(Status)*T	4	68.19	231	79.89	3.67E-16
c) Lethal temperature											
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)						
Null			245	2258.59							
Body size	1	927.17	244	1331.42	1.10E-15						
Status	2	1013.09	242	318.33	1.10E-15						
T	1	5.10	241	313.24	0.0009						
Species(Status)	4	216.27	237	96.97	1.10E-15						
Status*T	2	4.32	235	92.65	0.008						
Species(Status)*T	4	5.93	231	86.71	0.008						

Table 2. Results of the analysis of deviance for the effect of species identity (*Species*), acclimation temperature (*T*: 18, 24 and 30 °C), and their interaction on the upper (a-c) and lower (d-e) physiological thermal tolerance limits (CT_{max} and CT_{min} , respectively) of the species of the genus *Ophryotrocha* that survived the exposure to 30 °C (*O. labronica*, *O. japonica*, *O. adherens*). Degrees of freedom (Df), deviance residuals (Dev. Res.), degrees of freedom residuals (Df Res.), deviance (Dev.), and *P*-value with Benjamini-Hochberg correction for false positives (*P*(BH)) are provided

CT_{max}						CT_{min}					
a) Loss of locomotor control						d) Onset of spasms					
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)	Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)
Null			164	1547.8		Null			164	971.49	
Body size	1	108.61	163	1439.19	4.95E-16	Body size	1	71.39	163	900.10	4.95E-16
Species	2	550.98	161	888.20	4.95E-16	Species	2	52.99	161	847.11	4.95E-16
T	2	680.52	159	207.69	4.95E-16	T	2	717.89	159	129.22	4.95E-16
Species*T	4	108.62	155	99.07	4.95E-16	Species*T	4	56.28	155	72.94	4.95E-16
b) Onset of spasms						e) Chill coma					
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)	Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)
Null			164	446.19		Null			164	619.58	
Body size	1	41.85	163	404.34	2.2E-15	Body size	1	239.21	163	380.38	4.95E-16
Species	2	244.45	161	159.88	2.2E-15	Species	2	116.78	161	263.60	4.95E-16
T	2	74.4	159	85.49	2.2E-15	T	2	168.80	159	94.80	4.95E-16
Species*T	4	9.39	155	76.10	0.002	Species*T	4	32.35	155	62.44	4.95E-16
c) Lethal temperature											
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i> (BH)						
Null			164	626.63							
Body size	1	95.52	163	531.10	1.98E-15						
Species	2	462.47	161	68.63	1.98E-15						
T	2	23.64	159	44.99	1.98E-15						
Species*T	4	0.53	155	44.46	0.76						

Tab. 3. Results of the analysis of deviance for the effect of species identity (*Species*), acclimation temperature (*T*: 18, 24 and 30 °C) and their interaction on a) survival and b) total fecundity of five of the seven studied species of the genus *Ophryotrocha*. Degrees of freedom (Df), deviance residuals (Dev. Res.), degrees of freedom residuals (Df Res.), deviance (Dev.), and *P*-value with Benjamini-Hochberg correction for false positives (*P*(BH)) are provided

a) Survival					
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i>(BH)
Null			212	5541.00	
T	4	776.82	208	4764.20	1.1E-15
Species	2	2003.03	206	2761.20	1.1E-15
Species*T	8	2251.90	198	509.30	1.1E-15
b) Total fecundity					
Source	Df	Dev. Res.	Df Res.	Dev.	<i>P</i>(BH)
Null			212	107387	
T	4	25142	208	82245	1.1E-15
Species	2	28465	206	53780	1.1E-15
Species*T	8	26027	198	26027	1.1E-15

Figure legend

Fig. 1 Difference in (a-c) upper and (d-e) lower physiological thermal tolerance limits (CT_{max} and CT_{min} , respectively) among common (c, green border), uncommon (u, orange border) and rare (r, dark-red border) species of the genus *Ophryotrocha* following 7 d of exposure to 18 and 24 °C (white and grey fill, respectively). The boundary of the box closest to zero indicates the 25th percentile, a horizontal line and a black diamond symbol inside the box mark the median and mean, respectively, and the boundary of the box farthest from zero indicates the 75th percentile. Whiskers above and below the box indicate the 10th and 90th percentiles. Points above and below the whiskers indicate outliers. Capital letters indicate differences between species according to their common, uncommon, and rare status within acclimation temperature. An asterisk indicates differences between acclimation temperatures within status. Statistical significance: $P < 0.05$.
Sample size: $14 \leq N \leq 20$

Fig. 2 Difference in (a-c) upper and (d-e) lower physiological thermal tolerance limits (CT_{max} and CT_{min} , respectively) between species of the genus *Ophryotrocha* following 7 d of exposure to 18 and 24 °C (white and grey fill, respectively). Common (c), uncommon (u), and rare (r) species are indicated with green, orange, and dark red borders, respectively. See Fig. 1 for the description of the box plots. Capital letters indicate differences among species within acclimation temperature. An asterisk indicates differences between acclimation temperatures within status. Statistical significance: $P < 0.05$. Sample size: $14 \leq N \leq 20$. lab = *O. labronica*; jap = *O. japonica*; adh = *O. adherens*; puer = *O. puerilis*; hart = *O. hartmanni*; diad = *O. diadema*; rob = *O. robusta*

Fig. 3 Difference in a) survival and b) total fecundity between species of the genus *Ophryotrocha* measured during 90 d of exposure to 18, 24 and 30 °C (white, grey, and black fill, respectively). Common (c), uncommon (u), and rare (r) species are indicated with green, orange, and dark-red

borders, respectively. See Fig. 1 for the description of the box plots. Capital letters indicate differences among species within acclimation temperature. Lowercase letters indicate differences among acclimation temperatures within the same species. Statistical significance: $P < 0.05$. Sample size: $10 \leq N \leq 22$. lab = *O. labronica*; jap = *O. japonica*; adh = *O. adherens*; diad = *O. diadema*; rob = *O. robusta*

Fig. 4 Relationship between temperature and the habitat suitability index (HSI) defining the fundamental thermal niche of the common (green lines), uncommon (orange lines), and rare (dark red lines) species of the genus *Ophryotrocha*. Curves were interpolated across different levels of physiological impairment (CT_{min} and CT_{max}) and reproductive *optima*. CC = chill coma; LT = lethal temperature; OS = onset of spasms (CT_{min}); LLC = loss of locomotor control; $T_{optimal}$ = optimal temperature for reproduction. Subscripts *min*, *mean* and *max* refer to the minimum, mean or maximum values among the means calculated at each acclimation temperature for the traits considered

Fig. 5 Maps of the difference in the thermal habitat suitability (HSI) between the mid- (2050-2059) and late- (2090-2099) 21st Century scenario, and the present day (1970-2000) scenario under the RCP8.5 scenario in the North Atlantic and Mediterranean regions, obtained for the common (c), uncommon (u), and rare (r) species of the genus *Ophryotrocha*. Black arrows indicate areas characterized by HSI decreases

Figure 1.

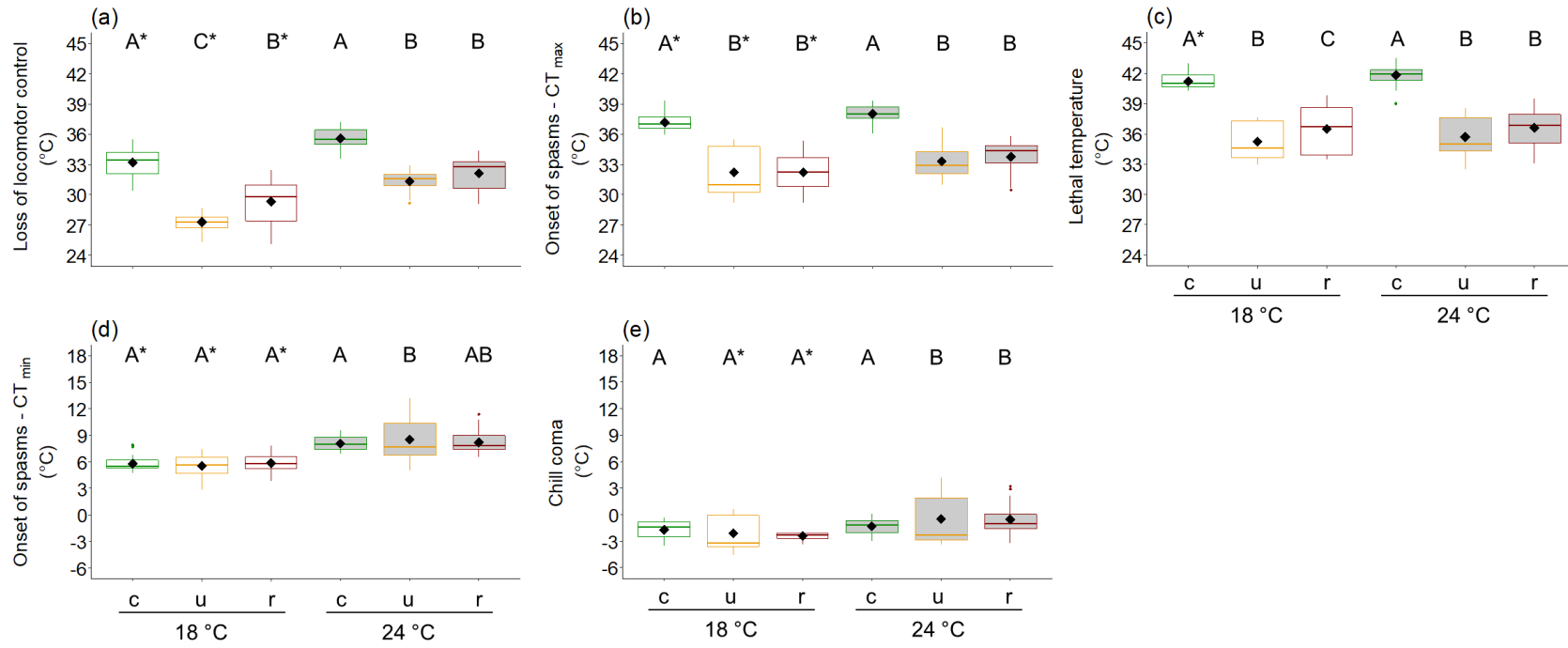


Figure 2.

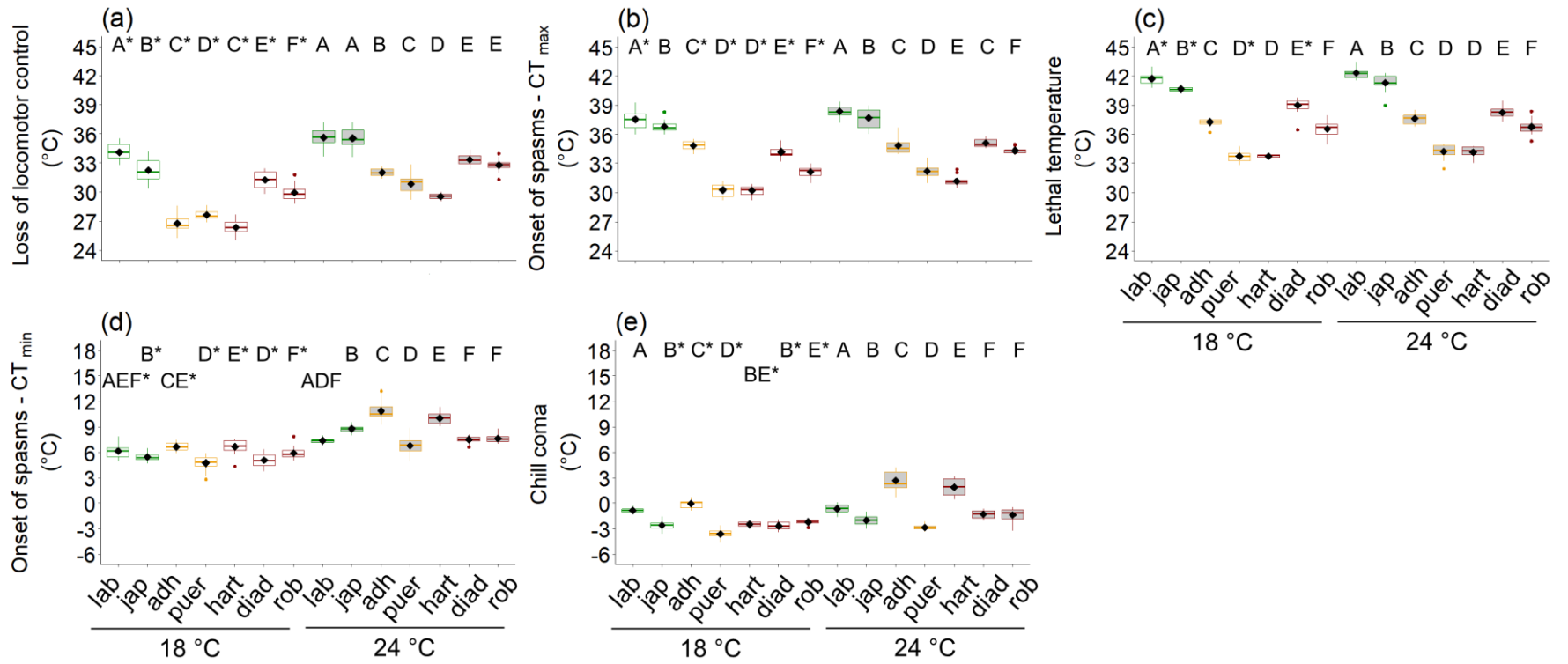


Figure 3.

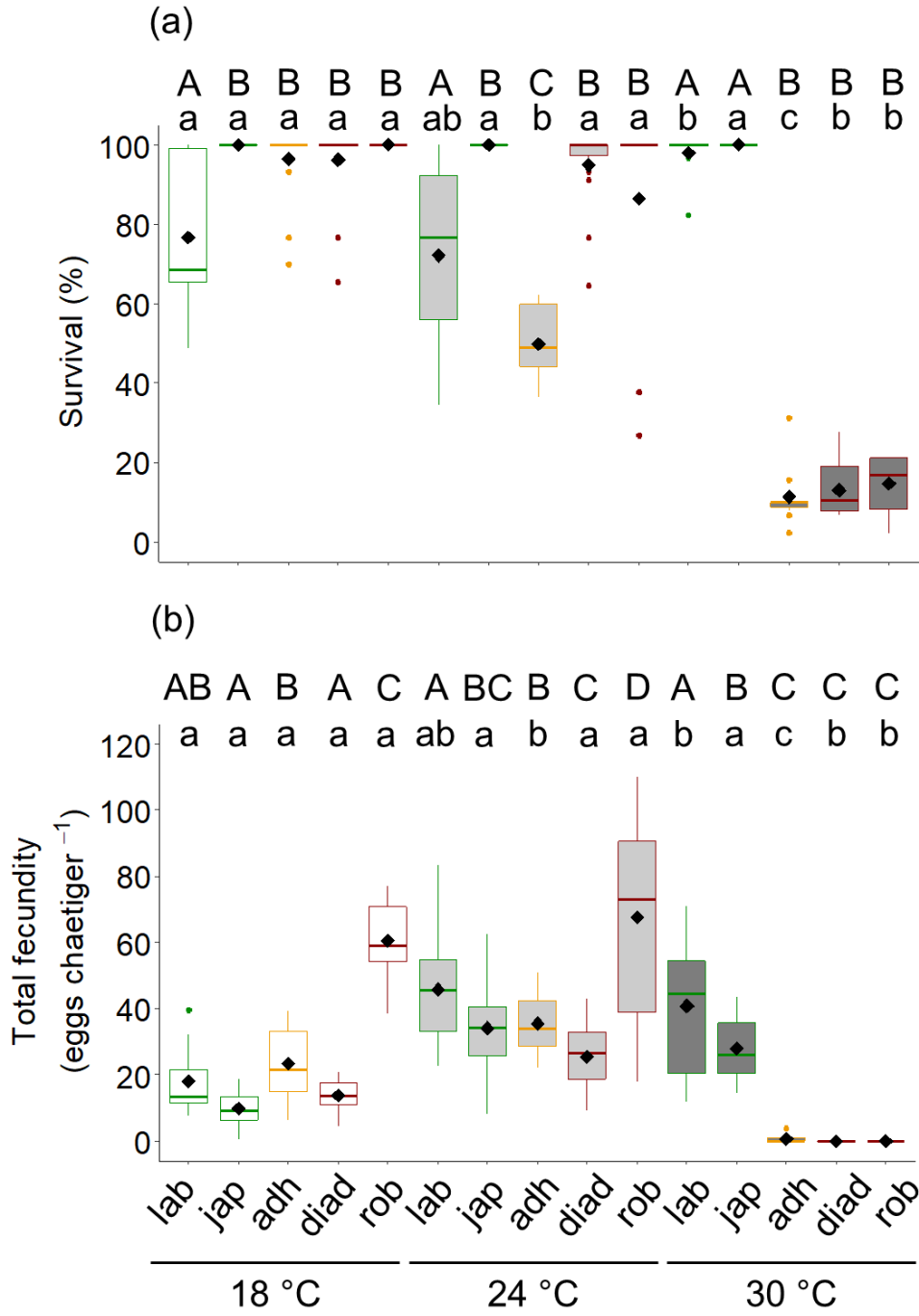


Figure 4.

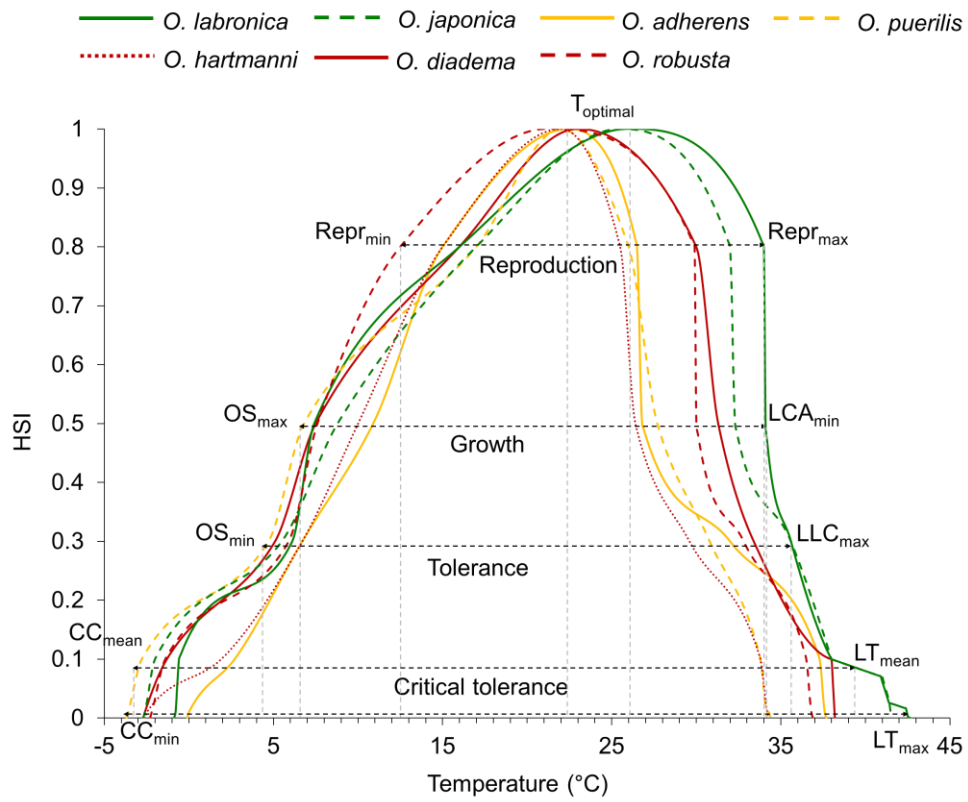


Figure 5.

